

1 Life cycle assessment of a residential building for environmental impact reduction in
2 Malaysia

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11
12 **Abstract**

13 The building industry has a significant impact on the environment due to massive natural
14 resources and energy it uses throughout its life cycle. This paper presents a typical residential
15 building construction in Malaysia as a case study and assesses the environmental impact under
16 cradle-to-grave which consist of pre-use, construction, use and end-of-life phase by using CML
17 2001. Five impact categories were evaluated namely, acidification, eutrophication, global
18 warming potential, ozone layer depletion and human toxicity. The building operation under use
19 phase contributed the highest global warming potential and acidification with 2.41E+03 kg
20 CO₂ eq and 1.10E+01 kg SO₂ eq respectively. In pre-use phase, concrete in the substructure
21 has the most significant impact overall especially to global warming potential with 3.40.E+02
22 kg CO₂ eq with cement as the primary raw materials. The recycling of steel and aluminium
23 during end-of-life shows a significant reduction by up to 78% in every environmental impact
24 category. The results showed that the residential building in Malaysia has a higher impact in
25 GWP and HT and lower in acidification and ODP compared to other studies. The results from
26 this study can assist policymakers and building professional to make informed decisions in

27 reducing environmental impact in Malaysia other countries in the world with similar
28 environmental condition.

29

30 **1 Introduction**

31 Building industry contributed significantly to the economy and social development, but
32 also responsible for the massive impact on the environment because of natural resource
33 consumption and emission released (Arena and de Rosa, 2003). Roodman et al. (1995)
34 suggested that buildings responsible for world's fresh water withdrawals, wood harvest and
35 material and energy flows that consist of 17%, 25% and 40% respectively. Due to the
36 significant environmental impact to the industry, numerous studies have been conducted to
37 reduce the energy consumption and its environmental impact (Singh et al., 2011).

38 Life cycle assessment (LCA) has been accepted as a tool to evaluate the environmental
39 impact throughout product life cycle (Fava et al., 2009; Oyarzo and Peuportier, 2014). The life
40 cycle of a product or cradle-to-grave which consist of the pre-use (extraction and acquisition
41 of raw materials, material production and manufacturing process), use and end-of-life (EOL)
42 was used to identify systematically and avoid the potential impact on the environment (ISO,
43 2006a). The introduction of LCA to building is relatively recent. The first study was conducted
44 by Adalberth (1997) that paved the research in this area. Recent reviews suggested that LCA
45 studies to buildings were conducted all over the world using ISO 14040 series as a basis, but
46 the methodologies were varied (Abd Rashid and Yusoff, 2015).

47 In Malaysia, LCA was initially introduced to assess the sustainability of palm oil
48 production (Ismail and Chen, 2010). Since then, it has been used in other industries such as
49 electronics, consumer goods, potable water production, electricity generation, waste
50 management and buildings (Bin Marsono and Balasbaneh, 2015; Fujita et al., 2008; Hassan et
51 al., 1999; Omar et al., 2014; Shafie et al., 2012; Sharaai et al., 2009a, 2009b; SIRIM, 2011;
52 Sumiani et al., 2009; Syafa Bakri et al., 2008; Thannimalay et al., 2013, 2012; Wen et al.,

53 2014). Buildings studies in Malaysia mainly focused on the impact assessment of different
54 materials. The studies also compared the benefit of integration of industrialised building system
55 (IBS) to conventional construction system. Fujita et al. (2008) used LCA to estimate CO₂
56 emission for concrete and timber based house using input-output method during pre-use and
57 operation phase. Omar et al. (2014) compared the pre-use phase of two-storey houses with a
58 conventional concrete house and an IBS system house with precast wall panel using hybrid
59 method for concrete and steel reinforcement. Wen et al. (2014) compared a conventional four-
60 storey apartments Johor Bahru and a four-storey IBS apartment in Iskandar Malaysia, Johor.
61 Bin Marsono and Balasbaneh (2015) compared seven different materials used for the wall of a
62 single-family unit house in Johor, but only global warming potential (GWP) was measured.

63 All studies mentioned were conducted without considering full building life cycle or
64 ‘cradle-to-grave’ which consist of pre-use, construction, use and end-of-life (EOL) phases.
65 Moreover, full environmental impact on residential buildings in Malaysia has yet to be
66 evaluated especially on the global warming impact. Thus, the aim of this study is to estimate
67 the life cycle impact of a residential building in Malaysia from cradle-to-grave in five impact
68 categories specifically on global warming potential (GWP), acidification, ozone depletion
69 (ODP) and eutrophication and human toxicity (HT). The results of this study provide detail
70 information which can be used to reduce the environmental impact of residential buildings in
71 Malaysia and to other locations in the world with similar environment condition.

72

73 **2 Methodology**

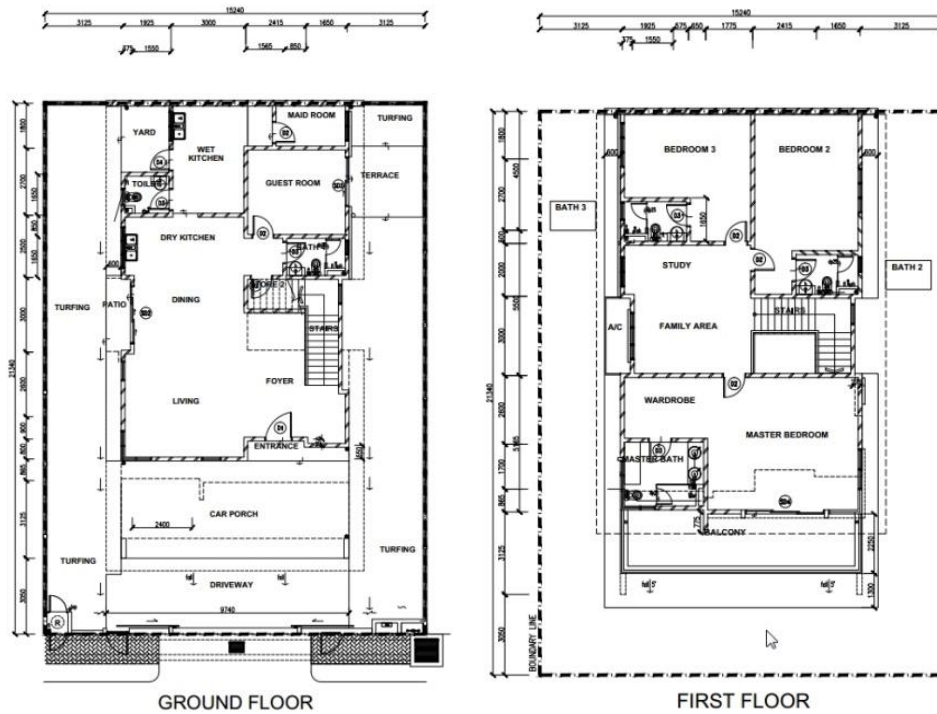
74 This study follows the LCA methodology standardised by ISO 14040 series which include
75 four stages namely goal and scope definition, life cycle inventory (LCI), life cycle impact
76 assessment (LCIA) and interpretation (ISO, 2006a, 2006b).

77

78 2.1 Goal and scope definition

79 The functional unit selected in this study was 1 m² of gross floor area (GFA), and the
80 building lifespan estimated within 50 years. The building type selected was a semi-detached
81 house within a residential development area located in the district of Seri Kembangan, Selangor
82 about 25 km from Kuala Lumpur. The building size is 246 m² GFA with five bedrooms, one
83 living room, a dining room, a dry kitchen, a wet kitchen, a family area, a study area and five
84 bathrooms as shown in Fig. 1. The building frame structures are reinforced concrete with clay
85 bricks as the building envelope.

86



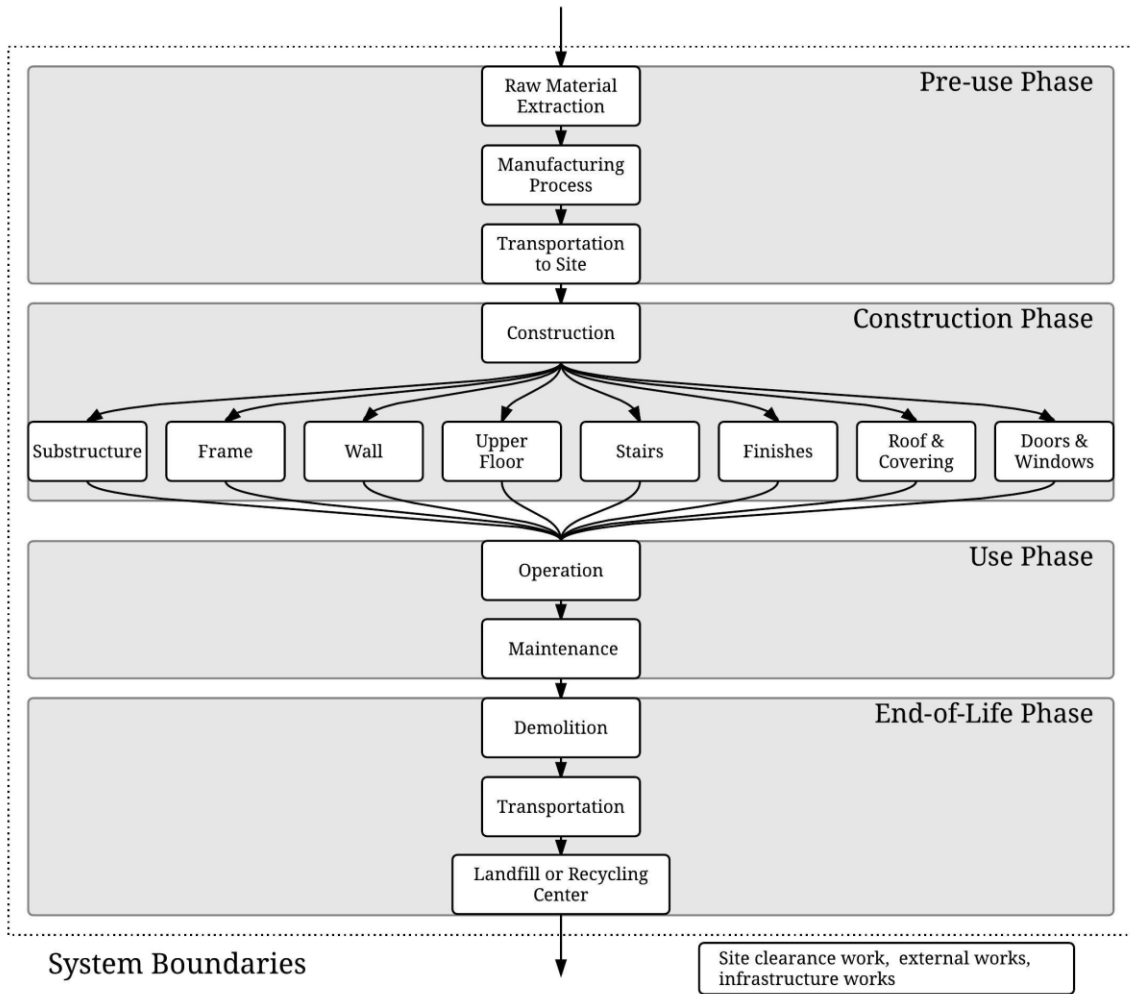
87

88 **Fig. 1** Floor plans of the house

89

90 2.1.1 System boundaries

91 The whole building life cycle was evaluated from cradle-to-grave within specific system
92 boundaries outlined in Fig. 2. The site clearance works, external works and infrastructure works
93 that involved the overall development were excluded which did not represent the case study.



94

95

Fig. 2 System boundaries of the life cycle model

96

97

The LCA modelling has been carried out in SimaPro V7.3.3 (PRé, 2015). Malaysia Life

98

Cycle Inventory Database (MYLCID) was used in the LCI especially on raw materials such as

99

cement and diesel to produce significant results for Malaysian scenario (MY-LCID, 2013). Due

100

to data limitation, Ecoinvent database was used and adapted to Malaysian conditions by

101

replacing the local electricity mix data set as suggested by Horváth & Szalay (2012).

102

103 2.2 *Life Cycle Inventory*

104 2.2.1 *Pre-use phase*

105 The data for LCI for the pre-use phase is obtained from the bill of quantities. The
106 quantities are then divided into GFA of the building as shown in Table 1. Few assumptions
107 have been considered due to the limitation of the databases as follows:

- 108 • An additional of 5% of material waste during construction was added as suggested by
109 previous studies (Buchan et al., 2003; Rossi et al., 2012).
- 110 • The types and materials are limited to process data equipped in the databases.
- 111 • Acrylic emulsion paint was substituted with alkyd paint due to the limitation in the
112 databases.
- 113 • The transportation distances from the manufacturer to the construction site were assumed
114 to be 300 km for all materials, meanwhile the distance is 50 km for ready-mix concrete, as
115 suggested by Wittstock et al. (2012).
- 116 • A 16-ton lorry was used to transport materials from manufacturers to site whereas a 24-ton
117 ready-mix lorry was used to transport concretes.

118 **Table 1**
119 LCI of materials in pre-use phase

Item	Materials	Quantity	Quantity/m ² GFA	Unit
A	Substructure			
	Excavation	86.02	0.35	m ³
	Hardcore	15.44	0.06	m ³
	Concrete Grade 7 blinding	21.74	0.09	m ³
	Concrete Grade 25	184.03	0.75	m ³
	Steel Reinforcement	2,561.62	10.41	kg
	Timber Formwork	4.13	0.02	m ³
B	Frame			
	Concrete Grade 25	23.20	0.09	m ³
	Steel Reinforcement	3,883.00	15.78	kg
	Timber Formwork	7.69	0.03	m ³
C	Upper Floor			
	Concrete Grade 25	28.73	0.12	m ³
	Steel Reinforcement	1,709.62	6.95	kg

Item	Materials	Quantity	Quantity/m ² GFA	Unit
D	Timber Formwork	3.35	0.01	m ³
	Stairs			
	Concrete Grade 25	2.78	0.01	m ³
	Steel Reinforcement	243.00	0.99	kg
E	Timber Formwork	1.07	0.00	m ³
	Brickwall			
	Clay brick			
F	Half brick thick	381.00	1.55	m ²
	One brick thick	37.14	0.15	m ²
	Roof and covering			
	Fascia board	0.31	0.00	m ³
	Painting	21.61	0.09	m ²
G	Timber Roof Trusses	10.65	0.04	m ³
	Clay roof coverings	213.84	0.87	m ²
	Finishes			
	Cement screed	9.47	0.04	m ³
	Ceramic tiles	357.59	1.45	m ²
	Timber strip	116.09	0.47	m ²
Plasterwork	18.57	0.08	m ³	
Painting	1,229.50	5.00	m ²	

120

121 2.2.2 Construction phase

122 Only three construction processes were taken into consideration namely excavation,
123 transportation of the excavator to the construction site and temporary timber formwork.
124 Excavator was used during excavation works meanwhile other installation works were assumed
125 to be completed by manual labours. The transportation of the excavator was considered to be
126 50 km distance from the construction site using a 40-ton low-loader. The formwork was
127 expected to be used multiple times before disposal as suggested by Abdullah (2005).

128

129 2.2.3 Use phase

130 2.2.3.1 Operation data

131 The total electricity consumption was estimated about 2,949.78 kWh per m² as shown in
132 Table 2. Energy simulation software OpenStudio V1.2.0 (OpenStudio, 2014) with EnergyPlus

133 was used to estimate the annual electricity consumption of air conditioning, illumination and
 134 electrical equipment. The Kuala Lumpur weather data for the year 2013 was used as the basis.
 135 The air-conditioning system was set at 20.8 degrees Celcius from 10.00 pm to 6.00 am every
 136 day in the master bedroom, and two other bedrooms on the first floor based on findings by
 137 Kubota et al. (2011). The electricity consumption was assumed to be constant throughout the
 138 operation of the house.

139

140 **Table 2**

141 Estimated electricity consumption during building operation for 50 years lifespan

142

Elements	Amount (kWh)
Air conditioning	341,236.50
Illumination	119,142.50
Electrical Equipment	265,266
Total	725,645.00
Total/GFA (kWh/m ²)	2,949.78

143

144 *2.2.3.2 Maintenance data*

145 Maintenance data was estimated based on selected elements such as painting,
 146 replacement of roof coverings and also changing of windows as suggested by other studies
 147 (Ochsendorf et al., 2011; Ortiz-Rodríguez et al., 2010). The replacement intervals are based on
 148 the report by National Association of Home Builders (NAHB) due to data limitations in
 149 Malaysia (Iyer-Raniga and Wong, 2012). The replacement interval is shown in Table 3, which
 150 includes the productions and transportation of selected building materials.

151

152 **Table 3**

153 Replacement interval of selected building elements in maintenance phase

Elements	Expected Lifespan	Number of replacement in 50 years
Painting	10 years	4 times
Roof covering	25 years	1 times
Window	30 years	1 times

154

155 2.2.4 *EOL phase*

156 EOL phase was incorporated into the LCA studies due to the ability of recycling
157 potential of building materials, which reduced the embodied energy (Blengini and Di Carlo,
158 2010). In Malaysia, only steel and aluminium were recycled whereas other materials are
159 transported to the landfill as suggested by Arham (2008). Recycled of aluminium and steel
160 scrap were used as the raw material instead of iron and aluminium ore to reduce environmental
161 impact as suggested by SimaPro (2015). The transportation distances from the construction site
162 to landfill and recycling centre were assumed to be 300 km.

163

164 2.3 *Life cycle impact assessment (LCIA)*

165 Only five common impact categories were used in this study for midpoints assessment
166 using CML 2001 namely, global warming potential (GWP), acidification, ozone depletion
167 (ODP) and eutrophication and human toxicity (HT) impact categories as suggested by
168 Khasreen et al. (2009).

169

170 2.4 *Interpretation*

171 LCIA will be interpreted according to the goal and scope of the study that shall include an
172 assessment and a sensitivity check of the significant inputs, outputs (ISO, 2006b). The findings
173 later will be validated by comparing it to other published studies (Ochsendorf et al., 2011).

174

175 **3 Results and discussion**

176 3.1 *Overview of the results*

177

178 Table 4 shows the result of the total LCIA. The operation phase within the building life
179 cycle has the highest impact on GWP and acidification compared to other phases due to fossil
180 fuel used in the electricity generation mix in Malaysia. The pre-use phase has the highest impact

181 on ODP and HT compared to other phases mainly contributed by materials and processes in
 182 the production of wall (6.28E-06 kg CFC-11 eq) and windows (8.66E+01 kg 1,4-DB eq)
 183 respectively. The construction phase has the lowest overall environmental impact which similar
 184 to previous studies (Blengini and Di Carlo, 2010; Ochsendorf et al., 2011; Rossi et al., 2012).
 185 EOL has the highest impact in eutrophication compared with other phases with 8.93E-01 kg
 186 PO4-eq which contributed by disposal of clay bricks to landfill. The maintenance phase has a
 187 lower impact in comparison to pre-use phase due to limited quantity of materials used.

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 189
 190

Table 4
 LCIA result based on building phases

Impact category	Unit	Total	Pre-use	Construction	Maintenance	Operations	End of Life
Acidification	kg SO2 eq	1.51E+01	3.23E+00	1.74E-02	4.09E-01	1.10E+01	4.62E-01
Eutrophication	kg PO4 eq	3.81E+00	6.19E-01	5.04E-03	1.49E-01	1.12E+00	1.92E+00
Global warming (GWP100)	kg CO2 eq	3.72E+03	8.25E+02	3.29E+00	9.19E+01	2.41E+03	3.89E+02
Ozone layer depletion (ODP)	kg CFC-11 eq	3.77E-05	2.39E-05	2.03E-07	9.42E-06	1.31E-07	4.05E-06
Human toxicity (HT)	kg 1,4-DB eq	7.77E+02	2.94E+02	1.56E+00	1.06E+02	2.36E+02	1.39E+02

191
 192

193 Table 5 shows the environmental impact of every element in pre-use phase. The
 194 substructure has the highest impact of acidification, eutrophication and GWP largely
 195 contributed due to the substantial quantity of cement in concrete based building elements which
 196 account for 77%, 53%, 81% respectively. The stair has the lowest overall impact due to small
 197 quantities of material use per GFA.

198
 199
 200

Table 5
 LCIA in pre-use phase based on building elements.

Impact category	Unit	Sub-structure	Frame	Wall	Upper floor	Stairs	Finishes	Roof & covering	Door	Window
Acidification	kg SO2 eq	1.47E+00	2.10E-01	3.31E-01	2.12E-01	2.24E-02	4.17E-01	6.99E-02	8.44E-02	1.53E-01
Eutrophication	kg PO4-eq	2.12E-01	5.00E-02	7.39E-02	2.80E-02	4.36E-03	9.13E-02	1.70E-02	3.31E-02	6.55E-02
Global warming (GWP100)	kg CO2 eq	3.48E+02	5.14E+01	1.09E+02	5.02E+01	5.41E+00	1.16E+02	2.05E+01	2.00E+01	4.01E+01

Ozone layer depletion (ODP)	kg CFC-11 eq	3.48E-06	1.07E-06	5.98E-06	4.16E-07	8.50E-08	4.96E-06	1.21E-06	1.68E-06	3.61E-06
Human toxicity	kg 1,4-DB eq	7.39E+01	1.66E+01	1.73E+01	9.96E+00	1.47E+00	3.26E+01	4.81E+00	3.51E+01	8.25E+01

201

202 During the maintenance phase, painting and aluminium frame window have been
 203 identified as the highest environmental impact contributors. Painting has the largest impact on
 204 acidification, eutrophication and ODP meanwhile, the aluminium frame window has the
 205 highest impact on GWP and HT as shown in Table 6.

206

207 **Table 6**
 208 LCIA in maintenance phase

Impact category	Unit	Aluminium window	Clay roof tiles	Painting
Acidification	kg SO2 eq	1.61E-01	3.54E-02	2.13E-01
Eutrophication	kg PO4-eq	6.88E-02	5.20E-03	7.52E-02
Global warming (GWP100)	kg CO2 eq	4.21E+01	1.46E+01	3.53E+01
Ozone layer depletion (ODP)	kg CFC-11 eq	3.79E-06	8.13E-07	4.81E-06
Human toxicity (HT)	kg 1,4-DB eq	8.66E+01	1.87E+00	1.78E+01

209

210 The environmental impact during EOL phase has the highest level of eutrophication with
 211 relatively high GWP. Table 7 shows the LCIA of transportation to the landfill and disposal of
 212 building materials. The impact of disposal of clay bricks was the highest in all impact categories
 213 followed by cement based products.

214

215 **Table 7**
 216 LCIA in EOL phase based on building materials

Impact category	Unit	Baseplaster	Cement Screed	Ceramic Tiles	Clay Roof Tiles	Concrete Blinding
Acidification	kg SO2 eq	1.51.E-02	7.79.E-03	8.99.E-03	9.14.E-03	1.33.E-02
Eutrophication	kg PO4 eq	3.04.E-01	1.58.E-01	1.82.E-01	1.82.E-01	3.57.E-03
Global warming (GWP100)	kg CO2 eq	5.44.E+0	2.81.E+0	3.24.E+0	3.25.E+0	2.60.E+0
Ozone layer depletion (ODP)	kg CFC 11 eq	4.12.E-07	2.13.E-07	2.46.E-07	2.47.E-07	1.13.E-07
Human toxicity	kg 1,4 DB eq	2.16.E+0	1.12.E+0	1.29.E+0	1.29.E+0	5.37.E-01

Impact category	Unit	Concrete G25	Door Single Glaze	Aluminium Door	Timber Door	Formwork
Acidification	kg SO2 eq	1.60.E-01	4.78.E-08	4.76.E-06	1.76.E-07	3.90.E-05
Eutrophication	kg PO4 eq	4.31.E-02	1.28.E-08	1.15.E-06	4.73.E-08	1.05.E-05
Global warming (GWP100)	kg CO2 eq	3.14.E+0	9.37.E-06	8.52.E-04	3.45.E-05	7.65.E-03
Ozone layer depletion (ODP)	kg CFC 11 eq	1.36.E-06	4.06.E-13	1.31.E-10	1.49.E-12	3.31.E-10
Human toxicity	kg 1,4 DB eq	6.48.E+0	1.93.E-06	2.04.E-04	7.11.E-06	1.58.E-03

Impact category	Unit	Clay Brick	Painting	Timber Product	Timber Strip	Aluminium Window
Acidification	kg SO2 eq	2.08.E-01	1.75.E-03	1.63.E-03	2.69.E-03	2.63.E-05
Eutrophication	kg PO4 eq	8.93.E-01	6.12.E-02	4.37.E-04	6.41.E-04	6.35.E-06
Global warming (GWP100)	kg CO2 eq	1.84.E+0	1.45.E+0	3.19.E-01	5.15.E-01	4.71.E-03
Ozone layer depletion (ODP)	kg CFC 11 eq	2	0	1.38.E-08	1.13.E-09	7.24.E-10
Human toxicity	kg 1,4 DB eq	6.16.E+0	4.98.E+0	6.58.E-02	1.94.E-02	1.13.E-03

217

218

219

220 3.2 Process contribution analysis

221 3.2.1 Impact of materials

222 Contribution analysis process has been carried out to identify materials or processes that
223 produce the highest impact on the environment. In pre-use phase, the substructure has been
224 identified as the largest impact for acidification, eutrophication and GWP. Concrete was
225 dominant in every impact category as shown in Table 8. In the production of concrete, cement
226 has been identified as the highest contributor, followed by transportation of concrete to site as
227 shown in Table 9.

228 **Table 8**

229 Process contribution of LCIA of substructure

Impact category	Unit	Hardcore (crushed stone)	Concrete	Steel Reinforcement
Acidification	kg SO2 eq	2.44.E-03	1.44.E+00	2.62.E-02
Eutrophication	kg PO4 eq	8.61.E-04	1.93.E-01	1.81.E-02
Global warming Potential (GWP100)	kg CO2 eq	4.32.E-01	3.40.E+02	7.28.E+00
Ozone layer depletion (ODP)	kg CFC-11 eq	4.09.E-08	2.96.E-06	4.84.E-07

Human toxicity kg 1,4-DB eq 3.97.E-01 6.79.E+01 5.67.E+00

230

231 **Table 9**

232 **Process contribution analysis of LCIA of concrete in selected impact categories**

Impact category	Unit	Total	Cement		Transportation of concrete		Remaining process	
Acidification	kg SO2 eq	1.47.E+00	1.13.E+00	(77%)	1.74.E-01	(12%)	1.66.E-01	(11%)
Eutrophication	kg PO4 eq	2.12.E-01	1.11.E-01	(52%)	4.48.E-02	(21%)	5.62.E-02	(27%)
Global warming potential (GWP100)	kg CO2 eq	3.48.E+02	2.82.E+02	(81%)	3.40.E+01	(10%)	3.20.E+01	(9%)

233

234 In comparison to other building elements, wall contributed the highest ODP mainly
 235 from transportation of natural gas with 3.24E-06 kg CFC-11 eq (54%) and crude oil
 236 productions with 2.47E-06 kg CFC-11 eq (41%). Windows contributed the highest HT from
 237 chromium oxides flakes with 5.20E+01 kg 1,4-DB eq (60%) used in the production of
 238 aluminium windows.

239

240 *3.2.2 Impact of recycling potential*

241 Only steel and aluminium were considered to measure the recycling potential impact. A
 242 comparison analysis of two data processes were conducted between recycle and normal
 243 production of 1 kg of steel reinforcement and 1 m² of the aluminium window frame. The
 244 reduction of LCIA was very significant in both building materials as shown in Table 10,
 245 especially to acidification and GWP in steel and almost all impact categories in aluminium.

246

247 **Table 10**

248 **Estimated reduction of LCIA by recycling of steel and aluminium**

Impact category	Unit	Steel Reinforcement (1 kg)			Aluminium Window Frame (1 m ²)			Reduction	
		Recycle	Normal	Reduction	Recycle	Normal	Reduction		
Acidification	kg SO2 eq	2.12.E-03	5.09.E-03	2.97.E-03	58%	7.63.E-01	2.09.E+00	1.32.E+00	63%
Eutrophication	kg PO4--- eq	1.64.E-03	3.18.E-03	1.54.E-03	48%	3.28.E-01	7.90.E-01	4.62.E-01	58%
Global warming (GWP100)	kg CO2 eq	6.22.E-01	1.47.E+00	8.51.E-01	58%	2.01.E+02	4.77.E+02	2.76.E+02	58%
Ozone layer depletion (ODP)	kg CFC-11 eq	4.31.E-08	5.59.E-08	1.28.E-08	23%	1.81.E-05	3.35.E-05	1.54.E-05	46%
Human toxicity	kg 1,4-DB eq	5.29.E-01	8.68.E-01	3.40.E-01	39%	4.14.E+02	1.92.E+03	1.51.E+03	78%

249

250 3.3 Sensitivity analysis

251 This step had been conducted to determine the influence of assumption in this study
252 specifically on the transportation distances. The predetermined distances were 50 km for
253 concrete and 300 km for other materials. The standard deviation of $\pm 20\%$ is allocated for
254 transportation distance as suggested by Wen et al. (2014). Substructure element was used as
255 base case scenarios as it has the largest impact. The result shows that the transportation
256 distances were not give a significant impact overall with the maximum effect of 8.78% in ODP
257 while other impact categories were below 6% variances as shown in Table 11.

258

259 **Table 11**

260 LCIA with $\pm 20\%$ standard deviation for transportation distance for substructure

Impact category	Unit	Percentage
Acidification	kg SO ₂ eq	3.06%
Eutrophication	kg PO ₄ eq	5.51%
Global warming (GWP100)	kg CO ₂ eq	2.54%
Ozone layer depletion (ODP)	kg CFC-11 eq	8.78%
Human toxicity	kg 1,4-DB eq	1.85%

261

262 3.4 Comparison with other studies

263 Since no full LCA studies for Malaysian residential building available, therefore detail
264 comparison is not possible. The comparison of selected impact categories at different type of
265 residential building with other studies as shown in Table 12. The GWP of this study was
266 $8.02E+02$ kg CO₂ eq while the flat was much lower at $3.44+E02$ kg CO₂ eq and $2.98.E+02$ kg
267 CO₂ eq. The ratio of certain elements such as roof, wall, floor and ceiling were shared between
268 multiple units which reduced the impact per m² GFA. The specification and the quantity per
269 m² also contributed to the difference, for example, detail specification of the brick used in the
270 4-storey flat were not explicitly mentioned. In general, a low to medium cost houses uses
271 cheaper cement-based brick which also has lower energy used in production in comparison to
272 clay bricks which influenced the overall GWP (Utama and Gheewala, 2008).

273

274 **Table 12**275 Comparison of selected impact categories at different type of residential building with other
276 studies

Building	Impact categories					References
	Global warming (GWP100) Cradle-to-gate (kg CO2 eq)	Global warming (GWP100) Cradle-to-grave (kg CO2 eq)	Acidification (kg SO2 eq)	Ozone layer depletion (ODP) (kg CFC-11 eq)	Human toxicity (HT) (kg 1,4-DB eq)	
Flat (cast in situ)	3.44.E+02	-	-	-	-	Wen et al. (2014)
Flat (IBS)	2.98.E+02	-	-	-	-	Wen et al. (2014)
Semi-detached house	-	3.50.E+03	-	-	-	Cuéllar-Franca and Azapagic (2012)
Semi-detached house	-	2.43.E+03	1.85.E+01	1.17.E-04	7.18.E+02	Ortiz et al. (2009)
Semi-detached house	8.02.E+02	3.72.E+03	1.51.E+01	3.77.E-05	7.77.E+02	This study

277

278 The comparison of GWP impact category for cradle-to-grave between 2 semi-detached
279 houses is relatively comparable. Since the larger share of GWP is from use phase, the method
280 used in determining the energy consumptions, climates and impact from electricity generation
281 in different countries contributed to the variation in the results. The results for acidification and
282 HT were also primarily contributed by the use phase for both buildings. However, the ODP for
283 this study was lower in comparison to other building and largely contributed by the pre-use
284 phase rather than the use phase.

285

286 **4 Conclusion**

287 The operation during the use phase of the semi-detached house in Malaysia produced the
288 highest GWP and acidification. In pre-use phase concrete in the substructure was the largest
289 contributor to the overall impact with cement as the primary raw material. In the maintenance
290 phase, painting has the highest impact on acidification, eutrophication and ODP, while
291 aluminium window frame has the highest impact on GWP and HT. In EOL, material such as
292 clay brick and concrete has been identified as the largest impact on landfill. The recycling

293 potential of steel and aluminium shows a massive reduction in all impact categories. The
294 results in this study showed that the residential building in Malaysia produced higher impact
295 in GWP and HT and lower in acidification and ODP in compared to other studies. The findings
296 from this study can help policymakers and building professionals to assess the environmental
297 impact of building materials of complete life cycle in Malaysia or other countries in the world
298 with similar environment condition.